Nitrate addition inhibited methanogenesis in paddy soils under long-term managements

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ABSTRACT


Rice fields are a major source of atmospheric methane (CH\textsubscript{4}). Nitrate has been approved to inhibit CH\textsubscript{4} production from paddy soils, while fertilization as well as water management can also affect the methanogenesis. It is unknown whether nitrate addition might result in shifts in the methanogenesis and methanogens in paddy soils influenced by different practices. Six paddy soils of different fertilizer types and groundwater tables were collected from a long-term experiment site. CH\textsubscript{4} production rate and methanogenic archaeal abundance were determined with and without nitrate addition in the microcosm incubation. The structure of methanogenic archaeal community was analysed using the PCR-DGGE (polymerase chain reaction denaturing gradient gel electrophoresis) and pyrosequencing. The results showed that nitrate addition significantly decreased the CH\textsubscript{4} production rate and methanogenic archaeal abundance in all six paddy soils by 70–100% and 54–88%, respectively. The quantity, position and relative intensity of DGGE bands exhibited differences when nitrate was added. Nitrate suppressed the growth of methanogenic archaeal species affiliated to Methanosetaeaceae, unidentified Euryarchaeota, Thaumarchaeota and Methanosarcinaceae. The universal inhibition of nitrate addition on the methanogenesis and methanogens can be adopted as a practice of mitigating CH\textsubscript{4} emission in paddy soils under different fertilization and water managements.

Keywords: Oryza sativa L.; mineral fertilizer; organic manure; biomethanation; archaeal 16S rRNA gene

Rice fields have been recognized as a major source of atmospheric methane (CH\textsubscript{4}), with up to 15~20% of the total global CH\textsubscript{4} emissions to the atmosphere (Conrad 2002, Scheer et al. 2008). A positive balance between methanogenesis and methanotrophy leads to the CH\textsubscript{4} emission from paddy soils. Methanogens, belonging to the domain Archaea, have a limited trophic spectrum comprised of simple substrates, including \(\text{H}_2 + \text{CO}_2\) and acetate (Boone et al. 1993). Nitrate can act as alternative hydrogen sink and inhibits CH\textsubscript{4} production from anoxic soil, which has been widely observed in paddy soils (Lu et al. 2000, Banger et al. 2012). However, paddy fields of the rice harvest area are under various management practices, which also affects methanogenesis through their effects on methanogens. The inhibitive efficiency of

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Fertilization and water managements are the main cultural practices that affect the methanogenesis. In situ studies have shown that the type of fertilizers differs in CH₄ emission (Snyder et al. 2009, Linquist et al. 2012). Organic matter incorporation markedly increased CH₄ emission, while the effects of mineral fertilizers on CH₄ emission are complex and sometimes contradictory (Cai et al. 1997, Linquist et al. 2012). An increase of CH₄ emission was observed with urea applied in continuously flooded rice fields (Lindau and Bollich 1993). In China, mid-season aeration is one basic practice for raising rice yields and is widely adopted in rice cultivation (Cai et al. 1997). In situ studies report a significant decrease of CH₄ emission by rice fields that are drained one or several times during the crop cycle (Yagi et al. 1996, Hadi et al. 2010). In addition, the groundwater table controlling the soil O₂ concentration and redox potential is another non-negligible practice, which affects the methanogenesis and methanogens in paddy fields. Submerged soils maintained at different Eh values differed in methanogenesis (Kludze and DeLaune 1995). Combined with the application of organic fertilizer, an appropriate groundwater table could provide the optimum conditions for methanogenesis. As a competitor for H₂, the inhibition of nitrate on the methanogenesis and methanogens in paddy soils influenced by different practices has not been thoroughly interpreted.

Red soils are widely distributed throughout the tropical and subtropical areas of South China and account for 6.5% of the total arable land. In this context, a total of six red paddy soils under different fertilization and groundwater tables were collected. Nitrate was added in the subsamples of the soils followed by microcosm incubation. The rate of CH₄ production was measured and the abundance and community structure of methanogenic archaea were analysed. The aims of this study were to determine (1) the inhibitive efficiency of nitrate on the methanogenesis and (2) the shifts of nitrate in the methanogens developed under different long-term fertilization and water managements.

**MATERIAL AND METHODS**

**Site description and soil sampling.** Surface soil samples from six treatments were collected from a long-term fertilization station: (i) high-water-level (groundwater table –20 cm) mineral fertilizer (A); (ii) low-water-level (groundwater table –80 cm) mineral fertilizer (B); (iii) high-water-level normal amount of organic manure (C); (iv) low-water-level normal amount of organic manure (D); (v) high-water-level high amount of organic manure (E), and (vi) low-water-level high amount of organic manure (F). The fertilization station was established on a Ferralic Cambisol soil derived from the Quaternary red clay in 1982 on the campus of the Hunan Agricultural University, Hunan province, China. The site has a subtropical monsoon climate and a crop succession of early rice, late rice and winter fallow. The nitrogen (N) fertilizer was provided as urea at 150 kg/ha, P as single superphosphate at 33 kg/ha for early rice or 13 kg/ha for late rice, and K as KCl at 125 kg/ha. The seasonal applications of high amount of organic manure were provided as fresh *Lolium perenne* L. at 45 000 kg/ha for early rice and 25 000 kg/ha for late rice. Those of normal amount were half of the high amount, respectively.

There were six replicates for each treatment in the long-term experiment. Three cores from each replicate were randomly sampled to a 0–20 cm depth. Each sample was a composite of eighteen random cores collected from a single treatment and three soil samples were collected for each treatment. The fresh soil was mixed thoroughly, and sieved through 2 mm screens for incubation and further analysis. Basic physicochemical properties of the soil samples were listed in Table 1.

**Nitrate addition and incubation.** For each treatment, sieved fresh soil (equivalent to 10.0 g DW (dry weight)) was added in a 120 mL serum bottles. Distilled water was added to keep the soil/water ratio at 1:1. Nitrate (NO₃⁻-N 50 mg/kg in KNO₃ solution) was added into half of the bottles. Then the serum bottles were capped with butyl stoppers and aluminum caps, flushed with N₂ for 6 min, and incubated without shaking at 30°C in the dark. On each of the five different sampling dates, 36 bottles (6 treatments × 2 nitrate incubations × 3 replicates) were sacrificed. The bottles after incubating for 1 h were sampled as day 0.

**Chemical measurements.** Gas samples (20 mL) were collected from the headspace of the serum bottles using a pressure-lock syringe on the sampling dates. The concentrations of CH₄ was analysed using a gas chromatograph (GC, 7890A, Agilent
Technologies, Palo Alto, USA) equipped with a flame ionization detector (FID). The concentrations of N was extracted with 2.0 mol/L KCl and determined by a segmented flow analyser (Skalar SAN Plus, Skalar Inc., Breda, The Netherlands).

**Soil DNA extraction.** After incubation for 7 days, triplicate bottles from each treatment were destructively sampled and the fresh soil was stored at –80°C within 15 min. Total soil DNA was extracted from 0.5 g soil using a bead beating as described in the manufacturer's instructions (MP Biomedicals, Santa Ana, USA). Purity and quantity of DNA were determined using a Nanodrop Technologies ND-1000 UV-Vis Spectrophotometer (NanoDrop Technologies, Wilmington, USA). DNA extracts were stored at –20°C until use.

**Quantitative PCR.** The abundance of methanogenic archaeal 16S rRNA gene was quantified by qPCR on a CFX96 Optical Real-Time Detection System (Bio-Rad, USA) in three biological replicates and each with three technical replicates using primer pairs 1106F-GC/1378R (Feng et al. 2012). The reaction was performed in a 20 μL mixture containing 10 μL SYBR Premix Ex Taq (Takara, Shanghai, China), 0.5 μmol of each primer, and 1 μL of DNA template. The reaction conditions were as follows: 95°C for 3 min, 40 cycles of 95°C for 30 s, 57°C for 20 s and 72°C for 30 s and followed by a final extension at 72°C for 6 min. DGGE was performed using a Dcode Universal Mutation Detection System (Bio-Rad). PCR products (about 200 ng) were loaded onto an 8% polyacrylamide (acrylamide: bisacrylamide = 37.5:1) gel with a linear gradient of 30–70% (100% denaturant contains 7 mol/L urea and 40% formamide) denaturant for methanogenic archaeal 16S rRNA gene. The gel was electrophoresed at 100 V for 14 h with a constant temperature of 60°C.

**Cloning, sequencing, and phylogenetic analysis.** The dominant bands in the DGGE gel were excised and DNA was eluted by incubating each band in 40 μL of sterilized distilled water over night at 4°C. The re-amplified PCR products using primer pairs 1106F/1378R of the eluted DNA solutions were cloned using pEASY-T1 vector (TransGen Biotech, Beijing, China) according to the instructions of the manufacturer. Three clones containing the correct gene insert fragments from each DGGE band were sequenced (Genscript, Nanjing, China). Phylogenetic analysis was performed with MEGA 4.0 according to the protocol described previously (Tamura et al. 2007).

**Statistical analyses.** Analysis of variance (ANOVA) followed by the Tukey’s post hoc test were performed to assess the differences within the data sets. All data were analysed using a statistical package, SPSS 18.0 and P < 0.05 was considered to be statistically significant.

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Table 1. Basic physicochemical properties of paddy soils under different fertilization and water managements

<table>
<thead>
<tr>
<th></th>
<th>pH</th>
<th>Organic C (g/kg)</th>
<th>Total N (g/kg)</th>
<th>Available N (mg/kg)</th>
<th>NH₄⁺-N (mg/kg)</th>
<th>NO₃⁻-N (mg/kg)</th>
</tr>
</thead>
<tbody>
<tr>
<td>A</td>
<td>5.51 ± 0.04ᵃ</td>
<td>14.9 ± 1.27ᵇᵃ</td>
<td>1.41 ± 0.20ᶜ</td>
<td>93.5 ± 6.36ᵃ</td>
<td>8.23 ± 0.80ᵃ</td>
<td>1.20 ± 0.44ᵇ</td>
</tr>
<tr>
<td>B</td>
<td>5.50 ± 0.09ᵃ</td>
<td>14.1 ± 1.16ᵇ</td>
<td>1.26 ± 0.19ᵈ</td>
<td>99.6 ± 11.8ᵃ</td>
<td>10.1 ± 3.14ᵃ</td>
<td>1.98 ± 0.12ᵇ</td>
</tr>
<tr>
<td>C</td>
<td>5.51 ± 0.02ᵃ</td>
<td>19.4 ± 2.41ᵃ</td>
<td>1.58 ± 0.23ᵇ</td>
<td>107 ± 7.80ᵃ</td>
<td>11.2 ± 1.26ᵇ</td>
<td>1.30 ± 0.35ᵇ</td>
</tr>
<tr>
<td>D</td>
<td>5.54 ± 0.03ᵃ</td>
<td>18.0 ± 1.53ᵇ</td>
<td>1.60 ± 0.16ᵇ</td>
<td>116 ± 17.5ᵃ</td>
<td>10.1 ± 0.44ᵃ</td>
<td>4.18 ± 1.43ᵃ</td>
</tr>
<tr>
<td>E</td>
<td>5.48 ± 0.10ᵃ</td>
<td>19.7 ± 1.45ᵃ</td>
<td>1.64 ± 0.16ᵇ</td>
<td>115 ± 23.6ᵃ</td>
<td>10.4 ± 0.76ᵃ</td>
<td>1.77 ± 0.49ᵇ</td>
</tr>
<tr>
<td>F</td>
<td>5.46 ± 0.12ᵃ</td>
<td>19.5 ± 2.09ᵃ</td>
<td>1.77 ± 0.24ᵃ</td>
<td>104 ± 17.7ᵃ</td>
<td>10.7 ± 0.28ᵃ</td>
<td>3.80 ± 1.85ᵇ</td>
</tr>
</tbody>
</table>

¹Available nitrogen refers to the nitrogen easily absorbed and utilized by crops in soil, including ammonium, nitrate, amino, amide and some simple peptides and protein compounds. Mean ± standard error. Different letters within each column indicate significant differences in soils (P < 0.05, Tukey’s test).
RESULTS

**Methanogenesis in paddy soils under long-term fertilization and water managements.** The contents of organic C and N in collected paddy soils were shown in Table 1. The nitrate reduction rates of the paddy soils were from 11.91 ± 0.20 mg/kg/day to 14.33 ± 0.28 mg/kg/day. The NO$_3^-$-N contents of the soils all decreased to around 1 mg/kg within 3 days since nitrate added (data not shown).

Without nitrate addition, the CH$_4$ production rates of organic manure treatments (C, D, E and F) were significantly higher than the mineral fertilizer treatments (A and B) between fertilizer types (Figure 1). Between groundwater tables, only the difference in high amount of organic manure treatments (E and F) reached a significant level. The results of two-way ANOVA concluded that fertilization had significant ($P < 0.001$) while groundwater table had no significant ($P = 0.09$) impact on the CH$_4$ production rate of paddy soils. The results of Pearson’s correlation analysis showed that the CH$_4$ production rate had no significant correlation with soil pH, organic carbon, nitrogen contents or nitrate reduction rate (data not shown).

With nitrate addition, the CH$_4$ production rate all decreased compared with those without nitrate addition. For the treatments (D, E and F) possessing relatively fast CH$_4$ production rate, nitrate addition significantly inhibited the methanogenesis. In A and B, approximately no CH$_4$ was produced after nitrate addition during the incubation. Nitrate addition decreased the CH$_4$ production rate by 69.6–100%.

**Methanogens in paddy soils under long-term fertilization and water managements.** Without nitrate addition, the methanogenic archaeal 16S rRNA gene copies of soil C were significantly higher than the other treatments (Figure 2). The results of two-way ANOVA concluded that fertilization ($P < 0.001$) as well as groundwater table ($P < 0.05$) both had significant impacts on the methanogenic archaeal abundances. The results of Pearson’s correlation analysis showed that the methanogenic archaeal abundance had no significant correlation with soil basic physiochemical properties and nitrate reduction rate (data not shown). With nitrate addition, the methanogenic archaeal 16S rRNA gene copies of all the treatments were significantly lower than those without addition. Nitrate addition decreased the methanogenic archaeal abundance by 53.5–87.9%.

DGGE profile based on the methanogenic archaeal 16S rRNA gene was shown in Figure 3. The
quantity, position and relative intensity of bands had differences among the treatments. The band quantities of mineral fertilizer treatments were higher than those of organic manure treatments. For all treatments, nitrate addition decreased the quantities of DGGE bands.

A total of 36 DGGE bands were recovered and the contained methanogenic archaeal 16S rRNA genes were sequenced (Figure 4). Bands 3 and 11 were detected only in the treatments without nitrate addition, and affiliated to Methanosaetaceae and unidentified Euryarchaeota, respectively. The intensities of bands 33 and 34 were dramatically lower in the treatments with nitrate added than those without nitrate added. They belonged to Thaumarchaeota and Methanosarinaceae, respectively. The intensity of band 10 belonging to Rice Cluster I was higher with nitrate addition.

DISCUSSION

With nitrate addition, the CH$_4$ production rate was significantly inhibited in all six paddy soils. For mineral fertilizer treatments, CH$_4$ production rate was completely suppressed. For organic fertilizer treatments, the inhibitive efficiency of nitrate ranged from 70–93%. The inhibition can be efficient in soils with a wide range of methanogenesis rate. For high groundwater table, the CH$_4$ production rate was decreased on average about 83%, while the inhibitive efficiency was on average about 93% for low groundwater table. Nitrate would be a universal inhibitor for CH$_4$ production in paddy soils under different groundwater conditions. The application of nitrate might be adopted to mitigate the CH$_4$ emission in rice fields. Furthermore, nitrate would also increase the crop yields as an N fertilizer for the nitrate-preferred plants (Fageria and Baligar 2005).

The inhibitory effect of nitrate on methanogens could be explained by two mechanisms. On one hand, there would be a lack in H$_2$ for methanogenesis during the phase of nitrate reduction due to the substrate competition by the nitrate reducers (Roy and Conrad 1999). This effect can also be observed when closed-circuit MFCs were operated in paddy soils, causing the substrate competition with methanogens by the exoelectrogenic bacteria (Zhong et al. 2017). On the other hand, the denitrification intermediates (nitrite, NO and N$_2$O) could bring toxicity to the methanogens and hence depress their growth and activity (Zumft 1993, Choi et al. 2006).

Based on the PCR-DGGE analysis, nitrate addition decreased the abundance of methanogens affiliated to Methanosaetaceae, unidentified Euryarchaeota, Thaumarchaeota and Methanosarinaceae. Scheid et al. (2003) reported that the methanogenic archaeal community was altered by nitrate addition,
Figure 4. Phylogenetic relationships of methanogenic archaeal 16S rRNA gene sequences derived from different fertilization and water management of paddy soils with or without nitrate addition. The phylogenetic trees were constructed by 1000-fold bootstrap analysis using the neighbour-joining method and maximum composite likelihood model. Bootstrap values (%) are shown at branch points (more than 50%). Aquifex pyrophilus (M83548) was used as the out-group. The scale bars represent 0.05 substitutions per nucleotide.
significant for those utilizing H2. The archaeal groups suppressed would be the major contributors to the CH4 production in the paddy soil samples collected (Yuan and Lu 2009, Dubey et al. 2014, Hernández et al. 2017). A decrease in their abundance resulted in the falling of CH4 production rate. Besides, archaeal groups affiliated to Methanomicrobiales and Methanobacteriaceae were not significantly affected by the nitrate addition. Methanogens would be dormant or starved but can still survive in unfavourable conditions (Watanabe et al. 2007). Their activities of CH4 production may be suppressed by competition or toxicity and the DNA of inactive cells can also be detected. Nevertheless, the growth of one species belonging to Rice Cluster I was slightly stimulated by nitrate addition. Certain methanogenic archaea were resistant to environmental stress and toxicity (Yuan et al. 2009). It might be more favourable for their metabolism when other methanogenic groups are suppressed to be inactive.

REFERENCES


