

## Soil Structure after 18 Years of Long-term Different Tillage Systems and Fertilisation in Haplic Luvisol

VLADIMÍR ŠIMANSKÝ<sup>1\*</sup> and MARTIN LUKÁČ<sup>2</sup>

<sup>1</sup>*Department of Soil Science, Faculty of Agrobiology and Food Resources,  
Slovak University of Agriculture, Nitra, Slovak Republic;*

<sup>2</sup>*School of Agriculture, Policy and Development, University of Reading, Reading, UK*

\*Corresponding author: [Vladimir.Simansky@uniag.sk](mailto:Vladimir.Simansky@uniag.sk)

### Abstract

Šimanský V., Lukáč M. (2018): Soil structure after 18 years of long-term different tillage systems and fertilisation in Haplic Luvisol. *Soil & Water Res.*, 13: 140–149.

Soil structure is a key determinant of many soil environmental processes and is essential for supporting terrestrial ecosystem productivity. Management of arable soils plays a significant role in forming and maintaining their structure. Between 1994 and 2011, we studied the influence of soil tillage and fertilisation regimes on the stability of soil structure of loamy Haplic Luvisol in a replicated long-term field experiment in the Dolná Malánta locality (Slovakia). Soil samples were repeatedly collected from plots exposed to the following treatments: conventional tillage (CT) and minimum tillage (MT) combined with conventional (NPK) and crop residue-enhanced fertilisation (CR+NPK). MT resulted in an increase of critical soil organic matter content (St) by 7% in comparison with CT. Addition of crop residues and NPK fertilisers significantly increased St values (by 7%) in comparison with NPK-only treatments. Soil tillage and fertilisation did not have any significant impact on other parameters of soil structure such as dry sieving mean weight diameters (MWD), mean weight diameter of water-stable aggregates ( $MWD_{WSA}$ ), vulnerability coefficient (Kv), stability index of water-stable aggregates (Sw), index of crusting (Ic), contents of water-stable macro- ( $WSA_{ma}$ ) and micro-aggregates ( $WSA_{mi}$ ). Ic was correlated with organic matter content in all combinations of treatments. Surprisingly, humus quality did not interact with soil management practices to affect soil structure parameters. Higher sums of base cations, CEC and base saturation (Bs) were linked to higher Sw values, however higher values of hydrolytic acidity (Ha) resulted in lower aggregate stability in CT treatments. Higher content of  $K^+$  was responsible for higher values of  $MWD_{WSA}$  and MWD in CT. In MT, contents of  $Ca^{2+}$ ,  $Mg^{2+}$  and  $Na^+$  were significantly correlated with contents of  $WSA_{mi}$  and  $WSA_{ma}$ . Higher contents of  $Na^+$  negatively affected St values and positive correlations were detected between  $Ca^{2+}$ ,  $Mg^{2+}$  and  $Na^+$  and Ic in NPK treatments.

**Keywords:** different soil management; index of crusting; soil organic matter; soil structure; vulnerability coefficient; water-stable aggregates

Luvisols belong to a category of very fertile soils suitable for growing a wide range of crop plants, collectively they are the most commonly utilized soils in agricultural production in Slovakia (12.9% of agricultural land) and cover around of 317 360 ha (BIELEK 2014). Luvisols are often referred to as texturally differentiated soils and part of metamorphic soils (Russia), sols lessivés (France), Parabraunerden

(Germany), chromosols (Australia) and luvisols (Brazil). The US nomenclature had them formerly classified as Grey Brown Podzolic soils, but now they belong to Alfisols with high-activity clays. These soils cover an estimated about 500–600 million ha worldwide, mainly in temperate regions such as the East European Plain and parts of the West Siberian Plain in Russia, parts of Central Europe, the North-East

<https://doi.org/10.17221/38/2017-SWR>

of the USA, but are also found in the Mediterranean region and in southern Australia.

Many arable soils in the Danube Lowland (Slovakia) are prone to physical and chemical degradation due to their low organic matter content, unfavourable particle-size distribution and deteriorated physical state, most often resulting from a repeated application of incorrect soil management practices. Management of key soil properties such as soil structure is therefore a key consideration when reversing their degradation. Soil structure is a major soil property since it regulates soil functions such as water movement, water content, oxygenation and temperature (KODEŠOVÁ *et al.* 2015; NEIRA *et al.* 2015). Soil structure also greatly influences plant germination and root growth (TORMENA *et al.* 2016) and is a key determinant of soil quality (BALL & MUNKHOLM 2015). Soil structure can be modified by soil management practices which significantly influence aggregation and structural development of soils (BRONICK & LAL 2005). On the one hand, conventional tillage may enhance the disruption of physical properties (ŠIMANSKÝ *et al.* 2016a) included soil structure (BEARE *et al.* 1994; ŠIMANSKÝ *et al.* 2016b). On the other hand, conservation tillage and residue management are thought to represent viable options for enhancing soil organic carbon stabilization by improving soil aggregation (CHOUDHURY *et al.* 2014). WANG *et al.* (2015) indicate that intensive tillage has a twofold impact upon the soil: aggregate redistribution due to the soil transfer process and the mechanical breakage of macroaggregates. It is suggested that the use of reduced tillage is not only an effective method for diminishing soil erosion, but also a feasible strategy for improving soil structure on hill slopes. Crop residue retention is important for sequestering soil organic carbon (SOC), controlling soil erosion, and improving soil quality (BLANCO-CANQUI & LAL 2004). Repeated application of organic residues and fertilisation can alter soil physical properties (BALDOCK *et al.* 1994).

Long-term field experiments are essential for the assessment of tillage and fertilisation effects on soil fertility. They provide the best possible means to observe changes in soil properties (NEUGSCHWANDTNER *et al.* 2014). As mentioned above, soil aggregation can be improved by management practices such as reduction of agro-ecosystem disturbance, improvement of soil fertility, increase in organic inputs, enhancement of plant cover, and decrease in SOC decomposition rate (BRONICK & LAL 2005). The objectives of this study therefore were to (i) quantify

the extent to which tillage and manure treatments influence soil structure stability in a Haplic Luvisol and to (ii) determine the relationship between chemical properties, soil organic parameters and parameters of soil structure stability as driven by soil tillage regime and fertilisation.

## MATERIAL AND METHODS

This study is based on a long-term experiment established in 1994 at the experimental station of Slovak Agriculture University Nitra, in Dolná Malá locality (lat. 48°19'00"; lon. 18°09'00"). The soil at the site is classified as loamy Haplic Luvisol (WRB 2014) with soil texture determined as: 36% of sand, 49% of silt and 15% of clay. Soil carbon content at the start of the experiment was 12.9 g/kg, while the cation exchange capacity was 14.7 cmol(p<sup>+</sup>)/kg, base saturation percentage reached 92.6% and active pH of the soil was 6.96. The local climate is warm and very dry, with mean annual temperature of 9.8°C and rainfall of 573 mm.

The experimental field (187 × 44 m) is divided into five blocks (A, B, C, D and E), each 35 m wide. Each block was divided into nine rows of 3 m, with a protection belt of 3 m between individual zones. Every block was replicated four times, a schematic layout of the experimental field is shown in Figure 1. The field experiment had the following annual crop rotation: (1) red clover (*Trifolium pratense* L.), (2) pea (*Pisum sativum* L. subsp. *hortense* (Neitr.)), (3) winter wheat (*Triticum aestivum* L.), (4) maize (*Zea mays* L.), and (5) spring barley (*Hordeum vulgare* L.). A full-factorial experimental layout of tillage (3 levels) and fertilisation (3 levels) was applied to subplots within each replicate block. Tillage treatments were represented by conventional tillage (CT, tillage depth 0.22–0.25 m), reduced tillage (RT, cultivation depth 0.12–0.15 m) and minimum tillage (MT, operation depth 0.10–0.12 m). Fertilisation treatments were as follows: (1) 0 without fertilization, (2) CR+NPK – crop residues added together with NPK fertilizers, (3) NPK – with added NPK fertilizers. Plant residues were returned to the soil in CR+NPK variants. Fertilisation was applied annually with the mean dose reaching N 80 kg/ha, P (P<sub>2</sub>O<sub>5</sub>) 45 kg/ha and K (K<sub>2</sub>O) 72 kg/ha. Fertilizers used in the experiment were labelled as nitre ammonium with dolomite (LAV 27), potassium chloride (KCl) and triple superphosphate (Ca(H<sub>2</sub>PO<sub>4</sub>)<sub>2</sub>·H<sub>2</sub>O). The doses of NPK were calculated by balance method.

For the purpose of this study, soil samples were collected twice a year (spring and autumn) from the depth of 0–0.25 m in two types of soil tillage (CT and MT) and two variants of fertilisation (CR+NPK and NPK). In each combination of treatments, three different sampling locations were chosen randomly. Soil samples were collected with a corer and pooled to generate an average sample. Samples were air-dried in the laboratory at air temperature and standard soil analyses were used for determination of soil pH in H<sub>2</sub>O (1 : 2.5 – soil/distilled water) and KCl (1 : 2.5 – soil/1 M KCl) and sorption parameters such as hydrolytic acidity (Ha), sum of base cations (SBC) – included individual exchangeable cations (Ca<sup>2+</sup>, Mg<sup>2+</sup>, Na<sup>+</sup> and K<sup>+</sup>), cation exchange capacity (CEC), and base saturation (Bs) (HRIVŇÁKOVÁ *et al.* 2011). Soil organic carbon was determined according to the Tyurin method (DZIADOWIEC & GONET 1999) and the composition of humus substances (C<sub>HS</sub>), humic (C<sub>HA</sub>) and fulvic (C<sub>FA</sub>) acids was determined according to Belchikova and Kononova (DZIADOWIEC & GONET 1999). The absorbance of humus substances and humic acids was measured at 465 and 650 nm to calculate the colour quotient Q<sub>4/6</sub><sup>HS</sup> and Q<sub>4/6</sub><sup>HA</sup>. Soil samples for determination of soil structure parameters were collected with the aid of a spade to maintain the soil aggregation. In laboratory, large clods were gently broken up along natural

fracture lines, followed by air-drying at the laboratory temperature. Soil samples were sieved (dry and wet sieving) to the following seven size fractions: > 7, 7–5, 5–3, 3–2, 2–1, 1–0.5, 0.5–0.25 mm. Percentages of water-stable aggregates (WSA) were determined by the Bakshev method (VADJUNINA & KORCHAGINA 1986). The size fractions of WSA were as follows: > 5, 5–3, 3–2, 2–1, 1–0.5, 0.5–0.25 mm. The remaining material except for water-stable microaggregates was quantified in each sieve. The microaggregate fraction was calculated as the difference between the total weight of the soil sample and the sums of macroaggregates. Fractions of WSA larger than 0.25 mm (> 0.25 mm) were considered water-stable macroaggregates (WSA<sub>ma</sub>) and fractions smaller than 0.25 mm (< 0.25 mm) water-stable microaggregates (WSA<sub>mi</sub>). On the basis of dry and wet sieving samples, we then calculated values of mean weight diameters for dry sieving (MWD), mean weight diameter of water-stable aggregates (MWD<sub>WSA</sub>), vulnerability coefficient (Kv) by VALLA *et al.* (2000) as well as the stability index of water-stable aggregates (Sw) by Henin, index of crusting (Ic) (LAL & SHUKLA 2004) and critical soil organic matter content (St) according to PIERI (1991).

Statistical analyses were performed using Statgraphics Centurion XV.I programme (Statpoint Technologies, Inc., USA). A one-way ANOVA model and the

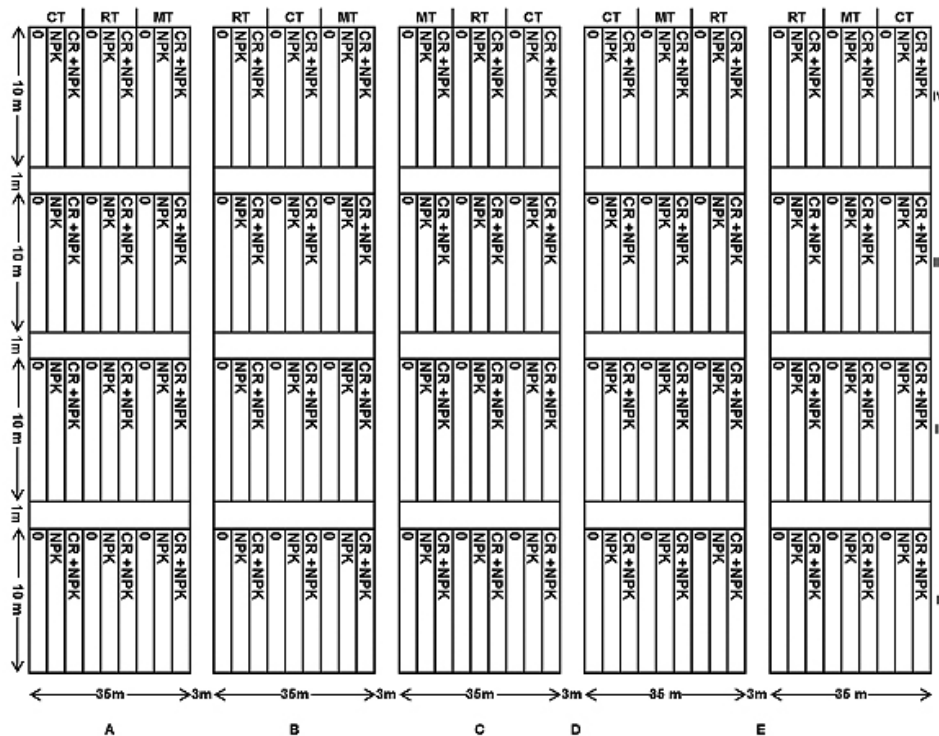


Figure 1. Schematic layout of the experimental field

<https://doi.org/10.17221/38/2017-SWR>

least significant difference tests were used to analyse the significance of differences in soil structure parameters between conventional and minimum tillage systems, as well as two levels of fertilisation. We used correlation analysis to determine the relationships between soil organic matter parameters, chemical properties and parameters of soil structure stability. Correlation coefficients were tested for significance at  $P < 0.05$ .

## RESULTS

**Soil structure parameters.** Parameters of soil structure affected by soil tillage and fertilisation are presented in Figure 2A–H. Soil tillage and fertilisation had a statistically significant effect on St, which increased by 7% in MT in comparison with CT. The incorporation of crop residues, together with NPK fertilizers (CR+NPK), significantly increased St values by 7% in comparison with NPK treatment. A better soil structure stability was achieved in MT than in CT, as indicated by the Kv values of 5.11 and 5.15,

respectively. A similar picture was revealed when looking at the presence and proportion of water-stable aggregates (Figure 2D, E). The mean value of Ic was 5% higher under NPK compared to CR+NPK, leading to higher Kv in the NPK treatment. Addition of just NPK fertilizers increased  $MWD_{WSA}$  and MWD by 9% and by 4% when compared to CR+NPK, respectively. The value of Sw (0.85) was slightly higher in CR+NPK than in NPK (0.82) treatments.

**Relationships between chemical properties and soil structure parameters.** We did not find any significant correlations between SOC and Kv,  $WSA_{mi}$ ,  $WSA_{ma}$ , Sw,  $MWD_{WSA}$  and MWD attributable to the application of tillage or fertiliser treatments. On the other hand, we found statistically significant correlations between SOC and St and Ic under all tillage and fertilisation regimes (Table 1). The content of  $C_{HS}$  negatively correlated with MWD only in MT and in NPK treatments. A higher content of  $C_{HA}$  affected  $WSA_{mi}$  content positively and  $WSA_{ma}$  content negatively under both CT and in CR+NPK

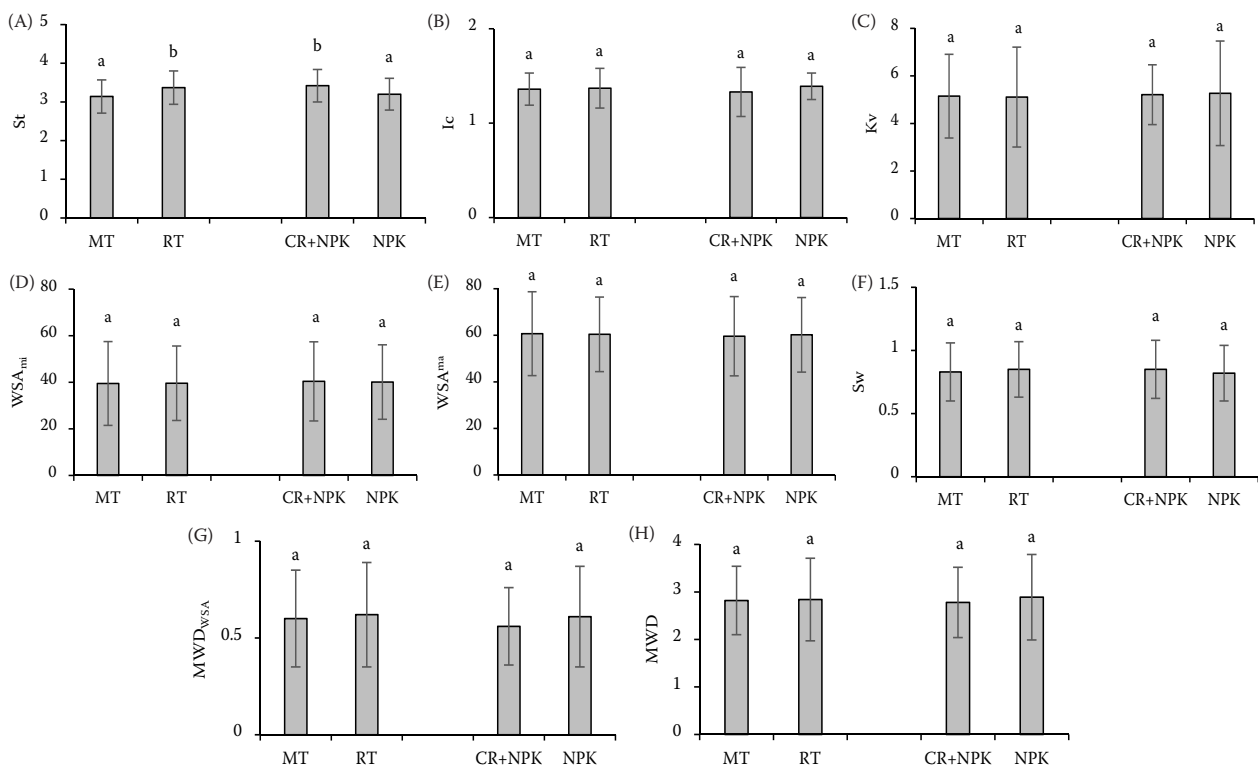


Figure 2. Analyses of variance of soil structure parameters: (A) critical level of soil organic matter, (B) index of crusting, (C) vulnerability coefficient, (D) content of water-stable microaggregates, (E) content of water-stable macroaggregates, (F) stability index of water-stable aggregates, (G) mean weight diameter of water-stable aggregates, (H) dry sieving mean weight diameters

Different letters between columns (a, b) indicate that treatment means are significantly different at  $P < 0.05$  according to the LSD multiple-range test

Table 1. Correlation coefficients between soil organic matter parameters and parameters of soil structure

	SOC	C <sub>HS</sub>	C <sub>HA</sub>	C <sub>FA</sub>	C <sub>HA</sub> :C <sub>FA</sub>	Q <sub>4/6</sub> <sup>HS</sup>	Q <sub>4/6</sub> <sup>HA</sup>
<b>Conventional tillage</b>							
St	0.672***	ns	ns	ns	ns	ns	ns
Ic	-0.377**	ns	ns	ns	ns	ns	ns
Kv	ns	ns	ns	ns	ns	ns	0.291*
WSA <sub>mi</sub>	ns	ns	0.314*	ns	ns	0.314*	ns
WSA <sub>ma</sub>	ns	ns	-0.311*	ns	ns	-0.324*	ns
Sw	ns	ns	ns	ns	ns	-0.314*	ns
MWD <sub>WSA</sub>	ns	ns	-0.294*	ns	ns	ns	ns
MWD	ns	ns	ns	ns	ns	ns	ns
<b>Minimum tillage</b>							
St	0.689***	ns	ns	ns	ns	ns	ns
Ic	-0.287*	ns	ns	ns	ns	ns	ns
Kv	ns	ns	ns	ns	ns	ns	ns
WSA <sub>mi</sub>	ns	ns	ns	ns	ns	ns	ns
WSA <sub>ma</sub>	ns	ns	ns	ns	ns	ns	ns
S <sub>w</sub>	ns	ns	ns	ns	ns	ns	ns
MWD <sub>WSA</sub>	ns	ns	ns	ns	ns	ns	ns
MWD	ns	-0.444***	-0.286*	-0.304*	ns	ns	ns
<b>NPK treatments</b>							
St	0.708***	ns	ns	ns	ns	ns	ns
Ic	-0.650***	ns	ns	ns	ns	ns	ns
Kv	ns	ns	ns	ns	ns	ns	ns
WSA <sub>mi</sub>	ns	ns	ns	ns	ns	ns	ns
WSA <sub>ma</sub>	ns	ns	ns	ns	ns	ns	ns
S <sub>w</sub>	ns	ns	ns	ns	ns	ns	ns
MWD <sub>WSA</sub>	ns	ns	ns	ns	ns	0.462**	ns
MWD	ns	-0.353*	ns	ns	ns	ns	ns
<b>CR+NPK treatments</b>							
St	0.666***	ns	ns	ns	ns	ns	ns
Ic	-0.688***	ns	ns	ns	ns	ns	ns
Kv	ns	ns	ns	ns	ns	ns	ns
WSA <sub>mi</sub>	ns	ns	0.356*	ns	ns	ns	ns
WSA <sub>ma</sub>	ns	ns	-0.356*	ns	ns	ns	ns
S <sub>w</sub>	ns	ns	ns	ns	ns	ns	ns
MWD <sub>WSA</sub>	ns	ns	-0.431**	ns	ns	ns	ns
MWD	ns	ns	-0.403*	ns	ns	ns	ns

ns – not significant; SOC – total soil organic carbon; C<sub>HS</sub> – humic substances; C<sub>HA</sub> – humic acids; C<sub>FA</sub> – fulvic acids; C<sub>HA</sub>:C<sub>FA</sub> – humic to fulvic acids ratio; Q<sub>4/6</sub><sup>HS</sup> – colour quotient of humic substances; Q<sub>4/6</sub><sup>HA</sup> – colour quotient of humic acids; St – critical level of soil organic matter; Ic – index of crusting; Kv – vulnerability coefficient; WSA<sub>mi</sub> – content of water-stable microaggregates; WSA<sub>ma</sub> – content of water-stable macroaggregates; Sw – stability index of water-stable aggregates; MWD<sub>WSA</sub> – mean weight diameter of water-stable aggregates; MWD – dry sieving mean weight diameters

<https://doi.org/10.17221/38/2017-SWR>

Table 2. Correlation coefficients between chemical properties and soil structure parameters

	Ha	SBC	CEC	BS	pH <sub>H<sub>2</sub>O</sub>	pH <sub>KCl</sub>	Ca <sup>2+</sup>	Mg <sup>2+</sup>	Na <sup>+</sup>	K <sup>+</sup>
<b>Conventional tillage</b>										
St	0.306*	-0.327**	-0.295*	-0.285*	-0.444***	-0.628***	-0.483***	-0.490***	-0.682***	ns
Ic	ns	ns	ns	ns	ns	0.279*	0.459***	0.429**	0.385**	ns
Kv	ns	ns	ns	ns	ns	ns	ns	-0.356**	ns	ns
WSA <sub>mi</sub>	0.479***	-0.424**	-0.358**	-0.467***	-0.426**	-0.509***	-0.271*	-0.372**	-0.393**	-0.393**
WSA <sub>ma</sub>	-0.482***	0.427**	0.360**	0.471***	0.434**	0.518***	ns	0.361**	0.409**	0.380**
Sw	-0.386**	0.348**	0.296*	0.380**	0.345**	0.353**	ns	ns	0.280*	0.349**
MWD <sub>WSA</sub>	-0.351**	ns	ns	0.310*	ns	0.269*	ns	0.339*	ns	0.284*
MWD	-0.292*	ns	ns	0.265*	0.356**	0.321*	ns	ns	ns	0.474***
<b>Minimum tillage</b>										
St	ns	ns	ns	ns	ns	ns	ns	ns	ns	ns
Ic	ns	ns	ns	ns	ns	ns	0.365**	0.377**	ns	ns
Kv	ns	ns	ns	ns	ns	ns	-0.283*	-0.403**	ns	ns
WSA <sub>mi</sub>	ns	ns	ns	ns	ns	ns	-0.300*	-0.380**	-0.322*	ns
WSA <sub>ma</sub>	ns	ns	ns	ns	ns	ns	0.300*	0.380**	0.322*	ns
Sw	ns	ns	ns	ns	ns	ns	ns	0.271*	ns	ns
MWD <sub>WSA</sub>	ns	ns	ns	ns	ns	ns	0.334*	0.352**	ns	ns
MWD	ns	ns	ns	ns	ns	ns	ns	ns	ns	ns
<b>NPK treatments</b>										
St	ns	ns	ns	ns	ns	ns	ns	ns	-0.387*	ns
Ic	ns	ns	ns	ns	ns	ns	0.438**	0.521**	0.374*	ns
Kv	ns	ns	ns	ns	ns	ns	ns	-0.432*	ns	ns
WSA <sub>mi</sub>	ns	ns	ns	-0.370*	ns	ns	ns	-0.423*	-0.357*	ns
WSA <sub>ma</sub>	ns	ns	ns	0.369*	ns	ns	ns	0.405*	0.383*	ns
Sw	ns	ns	ns	0.373*	ns	ns	ns	ns	ns	ns
MWD <sub>WSA</sub>	ns	ns	ns	ns	ns	ns	ns	0.476**	0.135	ns
MWD	ns	ns	ns	ns	ns	ns	ns	ns	ns	0.387*
<b>CR+NPK treatments</b>										
St	ns	ns	ns	ns	ns	ns	ns	ns	ns	0.430**
Ic	ns	ns	ns	ns	ns	ns	0.388*	ns	ns	ns
Kv	ns	ns	ns	ns	ns	ns	ns	ns	ns	ns
WSA <sub>mi</sub>	ns	ns	ns	ns	ns	ns	ns	ns	ns	-0.392*
WSA <sub>ma</sub>	ns	ns	ns	ns	ns	ns	ns	ns	ns	0.392*
Sw	ns	ns	ns	ns	0.340*	ns	ns	ns	ns	0.515**
MWD <sub>WSA</sub>	ns	ns	ns	ns	ns	ns	ns	ns	ns	0.427*
MWD	ns	ns	ns	ns	ns	ns	ns	ns	ns	0.481**

ns – not significant; Ha – hydrolytic acidity; SBC – sum of base cations; CEC – cation exchange capacity; BS – base saturation; St – critical level of soil organic matter; Ic – index of crusting; Kv – vulnerability coefficient; WSA<sub>mi</sub> – content of water-stable microaggregates; WSA<sub>ma</sub> – content of water-stable macroaggregates; Sw – stability index of water-stable aggregates; MWD<sub>WSA</sub> – mean weight diameter of water-stable aggregates; MWD – dry sieving mean weight diameters



treatments. Significantly negative correlations were found between  $C_{HA}$  and  $MWD_{WSA}$  in CT and between  $C_{HA}$  and  $MWD$  in MT. In comparison with NPK, we observed significantly negative correlations between  $C_{HA}$  and  $MWD_{WSA}$  and  $MWD$  in CR+NPK treatment. The degree of humification ( $C_{HA}:C_{FA}$ ) did not have any significant effects on soil structure parameters under any of the tested soil management practices. Higher humic substance stability supports the intensive aggregation of  $WSA_{mi}$  in CT, where it was also linked to a higher content of  $WSA_{ma}$  and higher aggregate stability. At the same time, higher presence of stable humic acids decreased  $K_v$  values. In MT however, no significant correlations between humic substance content and aggregate stability were observed. The same trend was observed in CR+NPK treatments. In NPK, the higher humic substance stability resulted in an increase of  $MWD_{WSA}$  values.

Under CT, we found several significant correlations between sorption parameters and soil structure parameters of Haplic Luvisol (Table 2). Sorption parameters and exchangeable cation content were strongly correlated with  $WSA_{mi}$  and  $WSA_{ma}$  content. Strong correlations were also observed between soil pH and the contents of exchangeable cations and  $St$  values. Higher contents of SBC, CEC and Bs resulted in higher  $Sw$  values, however higher values of Ha resulted in lower aggregate stability in CT treatments. Surprising positive correlations between  $Na^+$  and  $K^+$  and  $Sw$  were determined. In CT treatments, significantly negative correlations were also observed between Ha and  $MWD_{WSA}$  and  $MWD$  ( $r = -0.351$  and  $-0.291$ ,  $P \leq 0.01$  and  $P \leq 0.05$ , respectively). Both values of Bs and soil pH positively correlated with  $MWD_{WSA}$  and  $MWD$ . Higher content of  $K^+$  was responsible for higher values of  $MWD_{WSA}$  and  $MWD$  in CT.

However, correlations between sorption properties and soil structure parameters of Haplic Luvisol under RT indicated a different trend (Table 2). Several significant correlations between soil pH and sorption parameters and soil structure indicators driven by MT regime were observed.  $Ca^{2+}$  and  $Mg^{2+}$  were negatively correlated with  $K_v$  values in MT. The contents of  $Ca^{2+}$ ,  $Mg^{2+}$  and  $Na^+$  were also significantly correlated with  $WSA_{mi}$  and  $WSA_{ma}$  in MT, however a stronger correlation was found under CT (Table 2). On the other hand, more significant correlations between  $Ca^{2+}$  and  $Mg^{2+}$  and  $MWD_{WSA}$  than in CT were found under MT. As shown in Table 2, a few significant correlations were found between Ha as well as Bs values and  $WSA_{mi}$ ,  $WSA_{ma}$  and  $Sw$

in treatments with the addition of NPK fertilisers only. Higher contents of  $Na^+$  negatively affected  $St$  values. At the same time, we detected positive correlations between  $Ca^{2+}$ ,  $Mg^{2+}$  and  $Na^+$  and  $I_c$  in NPK treatment. In this treatment, higher content of  $Mg^{2+}$  and higher content of  $Na^+$  resulted in larger  $WSA_{ma}$  values (Table 2). A significant correlation was found between  $Mg^{2+}$  and  $MWD_{WSA}$  ( $r = 0.476$ ,  $P \leq 0.01$ ). In comparison with NPK treatments, some significant correlations between sorption parameters and soil structure parameters of Haplic Luvisol were determined in CR+NPK treatments. On the other hand, in comparison with NPK, in CR+NPK treatments the content of  $K^+$  correlated with  $St$ ,  $WSA_{mi}$ ,  $WSA_{ma}$ ,  $Sw$ ,  $MWD_{WSA}$  and  $MWD$  ( $r = 0.430$ ,  $-0.392$ ,  $0.392$ ,  $0.515$ ,  $0.427$ , respectively).

## DISCUSSION

Macroaggregate stability is often claimed to be affected by soil management practice (BRONICK & LAL 2005; BARTLOVÁ *et al.* 2015). Incorporation of crop residues into the soil is a very important tool for increasing macroaggregate stability (BALDOCK *et al.* 1994; BLANCO-CANQUI & LAL 2004). These observations were confirmed by this study, the addition of crop residues in combination with NPK application had a positive effect on  $St$  values after 18 years (Figure 2A). Under CR+NPK treatments, values of  $St$  were 7% higher than under NPK addition only. The application of mineral fertilisers may decrease soil organic matter content, which is often followed by lower aggregation. In our case, the likely reason for the decrease of SOM was the application of N fertiliser (LAV 27) in ammonium form which has been shown to negatively affect soil structure (HAYNES & NAIDU 1998). The higher ion concentration in the soil can result in higher clay dispersion and aggregate breakdown in fertilised soils (WHALEN & CHANG 2002).

The relationship between SOC and soil structure parameters was previously studied in different soil types, climate conditions and under varying soil management practices (e.g. BARTLOVÁ *et al.* 2015; RAJKAI *et al.* 2015; SCHACHT & MARSCHNER 2015), with contradictory results. For instance, Itami and Kyuma (1995) and Igwe *et al.* (1999) showed that SOC is not the most important binding agent responsible for aggregation. YILMAZ and SÖNMEZ (2017) mentioned a very strong linear relationship between SOC and  $MWD$ , an observation in direct

<https://doi.org/10.17221/38/2017-SWR>

contrast with our results (Table 1). We did not find any significant relationship between SOC and Kv,  $WSA_{mi}$ ,  $WSA_{ma}$ , Sw,  $MWD_{WSA}$  and MWD, nor did we see any modification of this relationship by soil tillage and fertilisation. However, we observed negative correlations between  $C_{HS}$  and MWD in MT and NPK treatments. In addition, we found statistically significant correlations between SOC and St and Ic under all soil management practices. The parameters of St and Ic are affected by particle-size distribution, SOC (PIERI 1991; LAL & SHUKLA 2004) and different soil management practices. Our results show that higher content of  $C_{HA}$  is related to WSA contents under CT and CR+NPK treatment. Strong correlations were detected between  $C_{HA}$  and  $MWD_{WSA}$  in CT and between  $C_{HA}$  and MWD in MT. In comparison with NPK, we observed negative correlations between  $C_{HA}$  and  $MWD_{WSA}$  and MWD in CR+NPK treatment. Surprisingly, some correlations were between the  $C_{HA}:C_{FA}$  ratio and soil structure parameters. OADES (1984) posits that not all organic compounds in the soil are responsible for aggregation. Different forms of organic matter stabilize aggregates of different sizes and sometimes may have no impact on the soil aggregation whatsoever or indeed may lower the aggregation potential. For example, the addition of anionic organic compounds (citrate, oxalate and acetate) to the soil can increase a dispersion of clay suspension (GOLDBERG *et al.* 1990; ITAMI & KYUMA 1995).

BOIX-FAYOX *et al.* (2001) claimed that larger aggregates are formed in soils with higher pH. This is in partial agreement with our results, as we found better indication of soil structure with increasing soil pH and SBC, but only under the CT treatment (Table 2). DIMOYIANNIS *et al.* (1998) connected aggregate stability with CEC, but NELSON *et al.* (1999) contended that this connection drives aggregate dispersion. In our case, higher values of CEC improved soil aggregate stability ( $r = 0.296$ ;  $P \leq 0.05$ ) in CT treatments.  $Ca^{2+}$ ,  $Mg^{2+}$  and  $Na^{+}$  contents were significantly correlated with the contents of  $WSA_{mi}$  and  $WSA_{ma}$ , however, a higher correlation was found in CT when compared to MT (Table 2). In general,  $Ca^{2+}$  is more effective in improving the soil structure than  $Mg^{2+}$  (ZHANG & NORTON 2002),  $Na^{+}$  can affect clay dispersion (MARCHUK *et al.* 2012), with resulting destruction of soil aggregates. In this experiment, the proportion of the sum of base cations as  $Ca^{2+}$  was highest under CT and lowest under MT (ŠIMANSKÝ & TOBIAŠOVÁ 2012), therefore we identified stronger

correlations between  $Ca^{2+}$ ,  $Mg^{2+}$  and soil structure parameters in CT than in MT treatments. Our results show that a higher content of  $Na^{+}$  results in higher contents of  $WSA_{ma}$  and in lower contents of  $WSA_{mi}$  in CT, MT and NPK treatments. This effect has not been observed in CR+NPK. Effects of  $K^{+}$  on structure parameters in CT and CR+NPK treatments were positive when compared with MT and NPK treatments, respectively. This effect can be related with higher contents of  $K^{+}$  in CT and in CR+NPK, as published by ŠIMANSKÝ and TOBIAŠOVÁ (2012).

## CONCLUSION

Soil structure parameters vary in time and the attributes observed at any given time reflect the accumulation of interacting factors over time. We did not observe any significant changes in parameters indicative of soil structure under any soil management practices, with the exception of St values. The correlations between SOC and soil structure indicators however show that (1) MT rather than conventional tillage and (2) ploughed crop residues together with NPK fertilizers rather than NPK only treatment are better for the management of favourable soil structure of Haplic Luvisol. We saw higher contents of humic acids and humic substances in CT and treatments with incorporated crop residues with NPK fertilizers. Since these compounds stimulate soil aggregate stability and support the intensive aggregation of water-stable aggregates, it is likely that CT and crop residues+NPK treatments are beneficial. In CT, aggregation processes were supported by better sorption parameters of soil compared to other soil management practices.

**Acknowledgements.** This study was partially supported by the Cultural and Educational Grant Agency (KEGA), Project No. 014SPU-4/2016 and the Scientific Grant Agency (VEGA), Project No. 1/0136/17.

## References

- Baldock J.A., Aoyama M., Oades J.M., Susanto R.H., Grant C.D. (1994): Structural amelioration of a south Australian red-brown earth using calcium and organic amendments. *Australian Journal of Soil Research*, 32: 571–594.
- Ball B.C., Munkholm L.J. (2015): *Visual Soil Evaluation: Realising Potential Crop Production with Minimum Environmental Impact*. Wallingford, CABI.
- Bartlová J., Badalíková B., Pospíšilová L., Pokorný E., Šarapatka B. (2015): Water stability of soil aggregates



<https://doi.org/10.17221/38/2017-SWR>

- in different systems of tillage. *Soil and Water Research*, 10: 147–154.
- Beare M.H., Cabrera M.L., Hendrix P.F., Coleman D.C. (1994): Aggregate-protected and unprotected organic matter pools in conventional and no-tillage soils. *Soil Science Society of America Journal*, 58: 787–795.
- Bielek P. (2014): Compendium to Practically Oriented Soil Science. Nitra, SUA. (in Slovak)
- Blanco-Canqui H., Lal L. (2004): Mechanisms of carbon sequestration in soil aggregates. *Critical Reviews in Plant Sciences*, 23: 481–504.
- Boix-Fayos C., Calvo-Cases A., Imeson A.C., Soriano-Soto M.D. (2001): Influence of soil properties on the aggregation of some Mediterranean soils and the use of aggregate size and stability as land degradation indicators. *Catena*, 44: 47–67.
- Bronick C.J., Lal R. (2005): The soil structure and land management: a review. *Geoderma*, 124: 3–22.
- Choudhury S.G., Srivastava S., Singh R., Chaudhari S.K., Sharma D.K., Singh S.K., Sarkar D. (2014): Tillage and residue management effects on soil aggregation, organic carbon dynamics and yield attribute in rice–wheat cropping system under reclaimed sodic soil. *Soil & Tillage Research*, 136: 76–83.
- Dimoyiannis D.G., Tsadilas C.D., Valmis S. (1998): Factors affecting aggregate stability of Greek agricultural soils. *Communications in Soil Science and Plant Analysis*, 10: 1239–1251.
- Dziadowiec H., Gonet S.S. (1999): Methodical Guide-book for Soil Organic Matter Studies. Warszawa, Polish Society of Soil Science. (in Polish)
- Goldberg S., Kapoor B.S., Rhoades J.D. (1990): Effect of aluminium and iron oxides and organic matter on flocculation and dispersion of arid zone soils. *Soil Science*, 150: 588–593.
- Haynes R.J., Naidu R. (1998): Influence of lime, fertilizer and manure applications on soil organic matter content and soil physical conditions: a review. *Nutrient Cycling in Agroecosystems*, 51: 123–137.
- Hrivňáková K., Makovníková J., Barančíková G., Bezák P., Bezáková Z., Dodok R., Grečo V., Chlpík J., Kobza J., Lištjak M., Mališ J., Píš V., Schlosserová J., Slávik O., Styk J., Širáň M. (2011): Uniform Methods of Soil Analyses. Bratislava, VÚPOP. (in Slovak)
- Igwe C.A., Akamigbo E.R., Mbagwu J.S.C. (1999): Chemical and mineralogical properties of soils in southeastern Nigeria in relation to aggregate stability. *Geoderma*, 92: 111–123.
- Itami K., Kyuma K. (1995): Dispersion behaviour of soils from reclaimed lands with poor soil physical properties and their characteristics with special reference to clay dispersion. *Soil Science and Plant Nutrition*, 41: 45–54.
- Kodešová R., Němeček K., Žigová A., Nikodem A., Fér M. (2015): Using dye tracer for visualizing roots impact on soil structure and soil porous system. *Biologia*, 70: 1439–1443.
- Lal R., Shukla M.K. (2004): Principles of Soil Physics. New York, Marcel Dekker.
- Marchuk A., Rengasamy P., McNeill A., Kumar A. (2012): Nature of the clay-cation bond affects soil structure as verified by X-ray computed tomography. *Soil Research*, 50: 638–644.
- Neira J., Ortiz M., Morales L., Acevedo E. (2015): Oxygen diffusion in soils: understanding the factors and processes needed for modeling. *Chilean Journal of Agricultural Research*, 75: 35–44.
- Nelson P.N., Baldock J.A., Clarke P., Oades J.M., Churchman G.J. (1999): Dispersed clay and organic matter in soil: their nature and associations. *Australian Journal of Soil Research*, 37: 289–315.
- Neugschwandtner R.W., Liebhard P., Kaul H.P., Wegentristl H. (2014): Soil chemical properties as affected by tillage and crop rotation in a long-term field experiment. *Plant, Soil and Environment*, 60: 57–62.
- Oades J.M. (1984): Soil organic matter and structural stability: Mechanisms and implications for management. *Plant and Soil*, 76: 319–337.
- Pieri C. (1991): Fertility of Soils: A Future for Farming in the West African Savannah. Berlin, Springer-Verlag.
- Rajkai K., Tóth B., Barna G., Hernádi H., Kocsis M., Makó A. (2015): Particle-size and organic matter effects on structure and water retention of soils. *Biologia*, 70: 1456–1461.
- Schacht K., Marschner B. (2015): Treated wastewater irrigation effects on soil hydraulic conductivity and aggregate stability of loamy soils in Israel. *Journal of Hydrology and Hydromechanics*, 63: 47–54.
- Šimanský V., Tobiašová E. (2012): Organic matter and chemical properties in Haplic Luvisol as affected by tillage and fertilizers intensity. *Acta Fytotechnica et Zootechnica*, 15: 52–56.
- Šimanský V., Polláková N., Jonczak J. (2016a): Is better minimum than standard mouldboard ploughing tillage from viewpoint of the pore-size distribution and soil water retention characteristic changes? *Cercetari Agronomice in Moldova*, 167: 17–26.
- Šimanský V., Polláková N., Jonczak J., Jankowski M. (2016b): Which soil tillage is better in terms of the soil organic matter and soil structure changes? *Journal of Central European Agriculture*, 17: 391–401.
- Tormena C.A., Karlen D.L., Logsdon S., Cherubin M.R. (2016): Visual soil structure effects of tillage and corn stover harvest in Iowa. *Soil Science Society of America Journal*, 80: 720–726.

<https://doi.org/10.17221/38/2017-SWR>

- Vadjunina A.F., Korchagina Z.A. (1986): *Methods of Study of Soil Physical Properties*. Moscow, Agropromizdat. (in Russian)
- Valla M., Kozák J., Ondráček V. (2000): Vulnerability of aggregates separated from selected anthrosols developed on reclaimed dumpsites. *Rostlinná Výroba*, 46: 563–568.
- WRB (2014): *World Reference Base for Soil Resources 2014. International Soil Classification System for Naming Soils and Creating Legends for Soil Maps*. World Soil Resources Reports No. 106, F Rome, AO.
- Wang Y., Zhang J.H., Zhang Z.H. (2015): Influences of intensive tillage on water-stable aggregate distribution on a steep hillslope. *Soil & Tillage Research*, 151: 82–92.
- Whalen J.K., Chang C. (2002): Macroaggregate characteristics in cultivated soils after 25 annual manure applications. *Soil Science Society of America Journal*, 66: 1637–1647.
- Yilmaz E., Sönmez M. (2017): The role of organic/bio-fertilizer amendment on aggregate stability and organic carbon content in different aggregate scales. *Soil & Tillage Research*, 168: 118–124.
- Zhang X.C., Norton L.D. (2002): Effect of exchangeable Mg on saturated hydraulic conductivity, disaggregation and clay dispersion of disturbed soils. *Journal of Hydrology*, 260: 194–205.

Received for publication February 7, 2017

Accepted after corrections November 13, 2017

Published online January 22, 2018