

Effects of sulfate on cadmium uptake in wheat grown in paddy soil – pot experiment

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Abstract: Rice-wheat rotation is common in China. Cadmium (Cd) and sulfur (S) are added to rice fields through various activities. The sulfur amendment has been recommended to control the uptake of Cd in rice. However, the effect of S on Cd uptake in winter wheat cultivated in paddy soil is rarely reported. A greenhouse pot experiment including two Cd levels (0, 10 mg Cd/kg, as CdCl₂) combined with three S rates (0, 30, 60 mg S/kg, as Na₂SO₄) was performed to investigate the effect of S application on uptake and allocation of Cd in wheat cultivated in paddy soil. Cadmium concentrations in wheat grain significantly ($P < 0.05$) increased by 37% at 30 mg S/kg, and the percentage of Cd allocation to grain significantly ($P < 0.05$) increased by 7% at 60 mg S/kg compared with non-S addition treatment when wheat was grown in Cd-added soil. For the low Cd soil, a similar trend was seen, but Cd increases were insignificant for grain while significant ($P < 0.05$) for root at 60 mg S/kg. In conclusion, S fertiliser may promote Cd accumulation in wheat grain and should be considered when it is used for wheat in paddy soils.

Keywords: cadmium toxicity; *Triticum aestivum* L.; plant uptake; translocation; distribution

Both cadmium (Cd) and sulfur (S) are added to rice fields through various activities (Chen et al. 2015, Gao et al. 2018). One-third of paddy soils in the main rice-growing area in China is contaminated with Cd (Mu et al. 2019). The S content of paddy soil is usually higher than that of dryland soil with the same parent material (Cao et al. 2011), due to the slow decomposition of organic matter including S in the anaerobic environment (Vityakon et al. 2000).

The annual paddy rice-winter wheat rotation covers around 10% of the total rice planting area in China (Frolking et al. 2002). Application of S has been recommended for reducing Cd accumulation in rice

(Fan et al. 2010), while its effect on grains of wheat grown in paddy soils is still not clear.

The present results about S fertilising effects on Cd uptake in plants are conflicting. For instance, Khan et al. (2015) revealed that sulfate application could enhance Cd tolerance by promoting S assimilation in wheat. Complexation of free Cd²⁺ by phytochelatin synthesized from S-containing glutathione, and deposition of Cd-phytochelatin into vacuoles is one of the mechanisms in response to Cd stress in wheat (Khan et al. 2007). Similar results were also found in maize (Adhikari et al. 2018). On the contrary, McLaughlin et al. (1998) demonstrated that sulfate

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application significantly increased Cd solubility in soil, which resulted in increasing concentrations of Cd in shoots of Swiss chard. In addition, Cd uptake in spring wheat increased in a pot experiment with K_2SO_4 supply (Zhao et al. 2003). In all, we aim at investigating whether S application affects Cd accumulation in wheat grain, as well as Cd distribution in wheat plant tissues.

MATERIAL AND METHODS

Soil used. The topsoil (0–20 cm) in a rice field was collected in September 2016 from Anqing city, Anhui province, China, where wheat-rice rotation is popular. The moist soil was air-dried, thoroughly mixed, and sieved through a 2-mm mesh. The main physicochemical properties are as follows: pH_{H_2O} 5.92; pH_{KCl} 5.42; total and available Cd contents were 0.35 mg/kg and 0.20 mg/kg; total and available S was 251 mg/kg and 15.5 mg/kg; soil texture was silt loam.

Experimental design. A greenhouse pot experiment was performed. There were six treatments, including two Cd rates (0, 10 mg Cd/kg, as $CdCl_2$) combined with three S levels (0, 30, 60 mg S/kg, as Na_2SO_4). Each treatment had four replicates. In total, 4 kg air-dried soil was placed in each polyethylene pot (20 cm diameter at the top, 18 cm diameter at the bottom, 20 cm height). All pots received the basic nutrients: 200 mg N/kg urea, 150 mg P/kg (KH_2PO_4), and 189 mg K/kg (KH_2PO_4). Cadmium, S, and the fertilisers were added to the soil in the form of solids and mixed thoroughly by a mixer machine (YG-5KG, Jinjie, Shenzhen, China). Each pot was performed individually.

Plant cultivation and sampling. Soil moisture content was controlled at 20–22% (v/v) during the entire experiment, and monitored continuously by a combined soil moisture sensor-meter (TR-6, Shunkeda, Beijing, China). In each pot, 12 seeds of the wheat cultivar Huaimai33 (*Triticum aestivum* L., Chinese Academy of Agriculture Sciences) were sowed and thinned to 4 plants two weeks later. The wheat grew at temperatures ranging between 0°C and 20°C, and relative humidity of 30% to 50%. All pots were arranged randomly. At maturity, 203 days after sowing, the whole plant was removed from the soil, and divided into 6 parts: root, stem, old leaves, top 3 leaves, husk, and grain. Soils on the root surface were collected as rhizosphere soils by shaking; the other soils in the pot were homogenized thoroughly and referred to as bulk soils (Wieland et al. 2001).

All plant parts were rinsed with deionized water and dried to constant weight at 70°C for 48 h, and dry weights (DWs) recorded.

Analysis of plant samples. Dried plant samples were finely grinded into powder with a stainless steel miller (FW80, Yongguangming, Beijing, China), except grain, which was grinded by hand with an agate mortar. Plant samples were digested according to Matusiewicz et al. (1989) with mixed acid (0.2 g plant + 8 mL HNO_3 + 3 mL H_2O_2) in a microwave oven (Mar 6, CEM Corporation, North Carolina, USA). Cadmium concentrations in the digestate were determined by inductively coupled plasma mass spectrometry (ICP-MS, Perkin Elmer, ELAN DRC-e, Waltham, USA), and S concentrations by inductively coupled plasma optical emission spectrometry (ICP-OES, Perkin Elmer, Optima 5300 DV, Waltham, USA). A reagent blank and standard materials (Celery GBW10048 (GSB-26), National Certified Reference Materials) were included to verify the accuracy and precision of the analysis. The recovery of Cd and S were 90–110% and 80–91%, respectively.

Analysis of soil samples. After harvesting, soils were extracted by diethylenetriamine pentaacetic acid (DTPA: 0.05 mol/L DTPA, 0.1 mol/L triethanolamine and 0.01 mol/L $CaCl_2$ at pH 7.3) at a soil:solution ratio of 1:2 (g/mL) as plant available Cd, and by 0.01 mol/L $Ca(H_2PO_4)_2$ at a soil:solution ratio of 1:5 (g/mL) as plant available S (Lu 1999). Cadmium and S in the extracts were determined by ICP-MS and ICP-OES, respectively. Soil pH was determined in a 1:2.5 suspension of soil and water or 1 mol/L KCl. Soil total Cd was determined by ICP-OES after digestion by HCl- HNO_3 -HF (Lu 1999), and total S by ICP-OES after digestion by $Mg(NO_3)_2$ (Lu 1999).

Data analysis. Cadmium transfer factors were the ratios of Cd concentrations in aboveground tissues (grain, husk, top 3 leaves, old leaves, stem) to that in roots (Boussen et al. 2013). Analysis of variance followed by the least significant difference (LSD) test and Pearson correlation analysis was performed using Window-based SPSS 25 (SPSS Inc., Armonk, USA).

RESULTS

Wheat plant growth. Cadmium addition significantly ($P < 0.05$) decreased number of spikes and grain DWs by 8% and 15% at zero mg S/kg level, as well as root DWs by 29% at 30 mg S/kg level, husk DWs and straw DWs by 23% and 22% at 60 mg S/kg level (Table 1).

The effects of S application on wheat growth were related to Cd levels and differed among different plant tissues (Table 1). Compared with non-S addition treatment, application of 30 mg S/kg significantly ($P < 0.05$) increased straw DWs and root DWs by 21% and 8% in –Cd treatments. In +Cd treatments, plant height and grain DWs significantly ($P < 0.05$) decreased by 7% and 9% at 30 mg S/kg, as well as straw DWs by 21% at 60 mg S/kg compared with non-S addition treatment. On the contrary, in +Cd treatments, DWs of root, straw, and husk decreased linearly with the increasing rate of S application. Consequently, total plant DWs presented no significant difference among S levels in –Cd treatments, but significantly ($P < 0.05$) decreased linearly in +Cd treatments, i.e., by 7.6% and 13% at 30 and 60 mg S/kg levels, respectively.

Uptake, distribution, and transfer of cadmium in wheat. The concentrations of Cd in all plant tissues were significantly ($P < 0.001$) higher in +Cd treatments than in –Cd treatments irrespective of S application. For instance, the grain Cd concentrations in +Cd treatments were 28, 34 and 23 times higher than in –Cd treatments at 0, 30, and 60 mg S/kg levels, respectively. The corresponding values for roots were 55, 46, and 22 times higher (Table 4).

There were interaction effects ($P < 0.001$) of S and Cd applications on Cd concentrations in wheat grains and stems. In –Cd treatments, the grain Cd concentrations tended to increase with the increase

of S levels, and exceeded the acceptable value of Cd (0.1 mg/kg) set in the maximum safe contaminant concentration standard in food of China (GB 2762-2017) by 50, 130 and 170% at 0, 30, 60 mg S/kg levels, respectively. However, in +Cd treatments, concentrations of Cd in grains and stems reached their peaks at the 30 mg S/kg level, and grain Cd concentrations exceeded the acceptable value (GB 2762-2017) 57–78 times. In +Cd treatments, there was no significant difference in Cd concentrations in other tissues (husk, leaves, roots) among S levels. However, for the –Cd treatment, S application significantly ($P < 0.05$) decreased Cd concentrations in leaves and increased that in roots. In +Cd treatments, the sequence of Cd concentrations in plant tissues was: root > stem \approx old leaves \approx husk > top 3 leaves > grain; a similar sequence was seen for the –Cd series, but differences were smaller (Table 4).

With the application of S at 0, 30, 60 mg/kg levels, the mean of total Cd uptake by the whole plant was 23.7, 21.5, and 20.9 $\mu\text{g}/\text{pot}$ in –Cd treatments, and 620, 615, 489 $\mu\text{g}/\text{pot}$ in +Cd treatments, respectively. Cadmium was predominately distributed in roots and straws, which accounted for 29% and 44%, respectively, followed by grains (15%) and husks (11%). Sulfur application tended to increase Cd allocation in grains and husks either in –Cd treatments or in +Cd treatments (Figure 1). In addition, S application significantly ($P < 0.05$) decreased Cd transfer factors

Table 1. Effect of sulfur (S) application on plant heights, number of spikes, dry weights of grain, husk, straw, root and total plant of wheat grown in paddy soil with and without the addition of 10 mg Cd/kg

| S level (mg/kg) | Plant heights (cm) | | Number of spikes (/pot) | | Grain (g/pot) | | Husk (g/pot) | | Straw ^a (g/pot) | | Root (g/pot) | | Total plant (g/pot) | |
|------------------------------------|--------------------|--------------------|-------------------------|-----------------|---------------------|------------------------|----------------------|----------------------|----------------------------|---------------------|----------------------|----------------------|---------------------|------------------------|
| | –Cd | +Cd | –Cd | +Cd | –Cd | +Cd | –Cd | +Cd | –Cd | +Cd | –Cd | +Cd | –Cd | +Cd |
| 0 | 59 $\pm 2^a$ | 61 $\pm 1^b$ | 12 $\pm 1^a$ | 11 $\pm 1^a$ | 13.0 $\pm 1.5^a$ | 11.0 $\pm 0.7^{ab}$ | 6.51 $\pm 0.55^a$ | 5.86 $\pm 0.89^a$ | 17.0 $\pm 1.0^a$ | 18.3 $\pm 1.1^b$ | 5.93 $\pm 1.17^a$ | 5.84 $\pm 1.73^a$ | 42.4 $\pm 1.7^a$ | 41.0 $\pm 3.0^b$ |
| 30 | 60 $\pm 3^a$ | 57 $\pm 1^a$ | 12 $\pm 1^a$ | 10 $\pm 1^a$ | 11.8 $\pm 1.6^a$ | 10.6 $\pm 0.5^a$ | 6.50 $\pm 1.06^a$ | 5.49 $\pm 0.73^a$ | 20.5 $\pm 2.4^b$ | 17.3 $\pm 1.2^b$ | 6.39 $\pm 0.87^b$ | 4.51 $\pm 0.97^a$ | 45.2 $\pm 5.0^a$ | 37.9 $\pm 2.3^{ab}$ |
| 60 | 61 $\pm 3^a$ | 59 $\pm 4^{ab}$ | 12 $\pm 2^a$ | 10 $\pm 1^a$ | 12.2 $\pm 0.6^a$ | 12.0 $\pm 1.1^b$ | 6.49 $\pm 0.57^a$ | 5.00 $\pm 0.49^a$ | 18.7 $\pm 2.1^{ab}$ | 14.5 $\pm 1.4^a$ | 4.33 $\pm 0.66^a$ | 4.18 $\pm 0.57^a$ | 41.7 $\pm 2.3^a$ | 35.7 $\pm 0.8^a$ |
| Analysis of variance for Cd levels | | | | | | | | | | | | | | |
| 0 | ns | | * | | * | | ns | | ns | | ns | | ns | |
| 30 | ns | | ns | | ns | | ns | | ns | | * | | * | |
| 60 | ns | | ns | | ns | | ** | | * | | ns | | ** | |
| S \times Cd | ns | | ns | | ns | | ns | | ** | | ns | | ns | |

^aStraw = top 3 leaves + old leaves + stem. Values are mean \pm standard deviations. Values followed by different lower case letters within a column indicate significances at $P < 0.05$ (least significant difference) for S application levels. ** $P < 0.01$, * $P < 0.05$ indicate the difference between –Cd and +Cd treatments for the same S level; ns – non-significant difference

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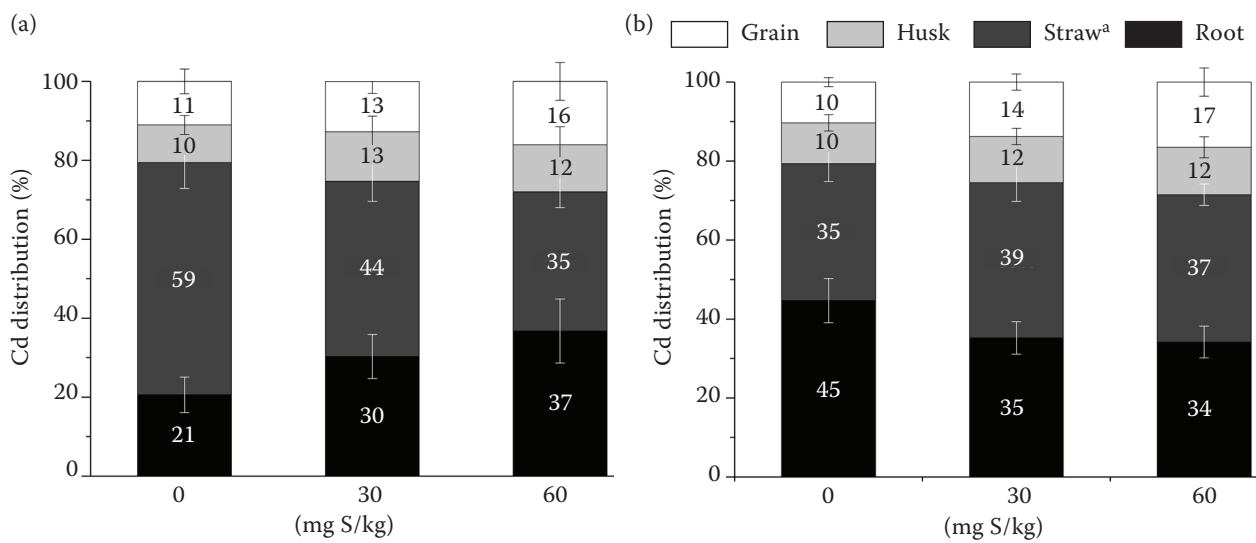


Figure 1. Effect of sulfur (S) application on the ratio of cadmium (Cd) distribution in plant tissues of wheat grown in paddy soil (a) without and (b) with addition of 10 mg Cd/kg. ^aStraw = top 3 leaves + old leaves + stem. Values demonstrate the mean ratio of Cd distribution in plant tissues. Error bars present standard deviations (*n* = 4)

of aboveground tissues with an exception for grain in –Cd treatments compared with non-S addition treatment, while tended to increase Cd transfer factors of aboveground tissues in +Cd treatments (Table 2).

Plant available Cd and S in soils. Generally, DTPA-extractable Cd in soils significantly (*P* < 0.001) increased by Cd addition, and it was slightly higher in rhizosphere soils than in bulk soils (Table 3). In addition, Ca(H₂PO₄)₂-extractable S in soil increased linearly (*P* < 0.001) with the increase of S application irrespective of Cd addition (Table 3).

DISCUSSION

In the present study, S and Cd application positively influenced Cd accumulation in wheat grains. In –Cd treatments, concentrations, and allocations of Cd in grains and roots tended to increase with S application (Table 4, Figure 1), as also seen from the decreasing Cd transfer factors from roots to aboveground tissues with increasing S application (Table 3). The results indicate that by wheat was cultivation in low Cd-contaminated paddy soils, S application

Table 2. Effect of sulfur (S) application on cadmium (Cd) transfer factors in above-ground tissues of wheat grown in paddy soil with and without the addition of 10 mg Cd/kg

| S level (mg/kg) | Grain | | Husk | | Top 3 leaves | | Old leaves | | Stem | |
|------------------------------------|-----------------------------|-----------------------------|------------------------------|-----------------------------|-----------------------------|-----------------------------|-----------------------------|-----------------------------|------------------------------|-----------------------------|
| | –Cd | +Cd | –Cd | +Cd | –Cd | +Cd | –Cd | +Cd | –Cd | +Cd |
| 0 | 0.25 ± 0.09 ^a | 0.13 ± 0.07 ^a | 0.44 ± 0.16 ^b | 0.23 ± 0.07 ^a | 0.82 ± 0.36 ^b | 0.13 ± 0.05 ^a | 1.59 ± 0.86 ^b | 0.29 ± 0.14 ^a | 0.70 ± 0.20 ^b | 0.25 ± 0.09 ^a |
| 30 | 0.24 ± 0.09 ^a | 0.17 ± 0.04 ^a | 0.41 ± 0.14 ^{ab} | 0.27 ± 0.06 ^a | 0.28 ± 0.13 ^a | 0.14 ± 0.03 ^a | 0.52 ± 0.19 ^a | 0.25 ± 0.07 ^a | 0.50 ± 0.25 ^{ab} | 0.34 ± 0.10 ^a |
| 60 | 0.16 ± 0.05 ^a | 0.17 ± 0.05 ^a | 0.22 ± 0.08 ^a | 0.30 ± 0.08 ^a | 0.15 ± 0.06 ^a | 0.19 ± 0.08 ^a | 0.15 ± 0.04 ^a | 0.32 ± 0.03 ^a | 0.29 ± 0.06 ^a | 0.33 ± 0.07 ^a |
| Analysis of variance for Cd levels | | | | | | | | | | |
| 0 | ns | | ns | | ** | | * | | ** | |
| 30 | ns | | ns | | ns | | * | | ns | |
| 60 | ns | | ns | | ns | | *** | | ns | |
| S × Cd | ns | | * | | *** | | ** | | * | |

Values are mean ± standard deviations. Values followed by different lower case letters within a column indicate significances at *P* < 0.05 (least significant difference) for S application levels. ****P* < 0.001; ***P* < 0.01; **P* < 0.05 indicate difference between –Cd and +Cd treatments at for the same S level; ns – non-significant difference

<https://doi.org/10.17221/558/2019-PSE>Table 3. Effect of sulfur (S) application on DTPA-extractable cadmium (Cd) and Ca(H₂PO₄)₂-extractable S (mg/kg) in the rhizosphere and bulk soils of wheat grown in paddy soil with and without the addition of 10 mg Cd/kg

| | S level (mg/kg) | DTPA-extractable Cd | | Ca(H ₂ PO ₄) ₂ -S | |
|----------------------|--------------------|---------------------|-------------|---|-------------|
| | | –Cd | +Cd | –Cd | +Cd |
| Rhizosphere soils | 0 | 0.10 ± 0.04 | 7.61 ± 0.35 | 20.6 ± 3.3 | 19.7 ± 5.8 |
| | 30 | 0.12 ± 0.02 | 8.03 ± 0.68 | 47.9 ± 22.8 | 39.9 ± 1.9 |
| | 60 | 0.15 ± 0.09 | 8.51 ± 0.35 | 74.5 ± 21.8 | 53.8 ± 18.8 |
| Bulk soils | 0 | 0.12 ± 0.03 | 7.66 ± 0.50 | 21.0 ± 2.7 | 17.7 ± 0.7 |
| | 30 | 0.12 ± 0.06 | 7.17 ± 0.71 | 44.0 ± 18.3 | 29.6 ± 1.6 |
| | 60 | 0.12 ± 0.07 | 7.09 ± 0.77 | 65.5 ± 9.8 | 80.4 ± 20.3 |
| Analysis of variance | | | | | |
| S | | | ns | | *** |
| Cd | | | *** | | ns |
| Distance (D) | | | ** | | ns |
| S × Cd | | | ns | | ns |
| S × D | | | * | | ns |
| Cd × D | | | ** | | ns |
| S × Cd × D | | | ns | | ns |

Values are mean ± standard deviations. *** $P < 0.001$; ** $P < 0.01$; * $P < 0.05$; ns – non-significant difference

reduced Cd translocation to aboveground tissues and thus caused higher root Cd contents. This agrees with the results reported by Khan et al. (2015) and Adhikari et al. (2018), who demonstrated that the production of glutathione and phytochelatin was stimulated by increasing S supply. The phytochelatin in wheat roots may restrain Cd and limit its translocation from root to shoot (Khan et al. 2007). The tendency of S facilitated

Cd uptake and accumulation in roots is also consistent with results observed by McLaughlin et al. (1998), for Swiss chard. The higher Cd uptake might be due to the increase of Cd phytoavailability in rhizosphere soil when sulfate is added (Table 3). Cadmium solubility in soil solution was observed to increase due to Cd-sulfate complexation by sulfate addition (McLaughlin et al. 1998). Furthermore, Cd uptake by *Zea mays* L. was

Table 4. Effect of sulfur (S) application on concentrations (mg/kg) of cadmium (Cd) in tissues of wheat grown in paddy soil with and without the addition of 10 mg Cd/kg

| S level (mg/kg) | Grain | | Husk | | Top 3 leaves | | Old leaves | | Stem | | Root | |
|------------------------------------|-----------------------------|-----------------------------|-----------------------------|----------------------------|-----------------------------|-----------------------------|-----------------------------|----------------------------|-----------------------------|----------------------------|-----------------------------|-----------------------------|
| | –Cd | +Cd | –Cd | +Cd | –Cd | +Cd | –Cd | +Cd | –Cd | +Cd | –Cd | +Cd |
| 0 | 0.20 ± 0.05 ^a | 5.78 ± 0.30 ^a | 0.34 ± 0.08 ^a | 10.9 ± 1.6 ^a | 0.64 ± 0.21 ^b | 6.16 ± 0.46 ^a | 1.21 ± 0.39 ^b | 13.2 ± 1.4 ^a | 0.56 ± 0.17 ^a | 11.5 ± 1.1 ^a | 0.91 ± 0.58 ^a | 51.2 ± 17.1 ^a |
| 30 | 0.23 ± 0.06 ^a | 7.94 ± 0.89 ^b | 0.40 ± 0.08 ^a | 13.0 ± 2.1 ^a | 0.26 ± 0.07 ^a | 6.69 ± 1.80 ^a | 0.53 ± 0.24 ^a | 11.9 ± 2.3 ^a | 0.47 ± 0.10 ^a | 16.1 ± 2.1 ^b | 1.05 ± 0.35 ^a | 49.2 ± 11.4 ^a |
| 60 | 0.27 ± 0.08 ^a | 6.55 ± 0.47 ^a | 0.37 ± 0.10 ^a | 11.5 ± 0.7 ^a | 0.26 ± 0.10 ^a | 7.38 ± 1.23 ^a | 0.26 ± 0.06 ^a | 13.3 ± 3.9 ^a | 0.50 ± 0.08 ^a | 12.9 ± 1.9 ^a | 1.76 ± 0.25 ^b | 40.9 ± 11.3 ^a |
| Analysis of variance for Cd levels | | | | | | | | | | | | |
| 0 | | *** | | *** | | *** | | *** | | *** | | *** |
| 30 | | *** | | *** | | *** | | *** | | *** | | *** |
| 60 | | *** | | *** | | *** | | *** | | *** | | *** |
| S × Cd | | *** | | ns | | ns | | ns | | ** | | ns |

Values are mean ± standard deviations. Values followed by different lower case letters within a column indicate significances at $P < 0.05$ (least significant difference) for S application levels. *** $P < 0.001$, ** $P < 0.01$ indicate the difference between –Cd and +Cd treatments at for the same S levels; ns – non-significant difference

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Table 5. The Pearson correlation coefficients between concentrations of plant-available cadmium (Cd) in soils and concentrations of Cd in plant tissues and among itself

| | | Concentrations of Cd in wheat plant tissues | | | | | |
|---|--------------|---|-------|--------------|------------|-------|-------|
| | | grain | husk | top 3 leaves | old leaves | stem | root |
| Concentrations of Cd in wheat plant tissues | husk | 0.979* | | | | | |
| | top 3 leaves | 0.945 | 0.929 | | | | |
| | old leaves | 0.929 | 0.947 | 0.910 | | | |
| | stem | 0.970 | 0.961 | 0.955 | 0.914 | | |
| | root | 0.895 | 0.948 | 0.964 | 0.926 | 0.922 | |
| Concentrations of DTPA-Cd in soils | rhizosphere | 0.981 | 0.978 | 0.959 | 0.960 | 0.963 | 0.914 |
| | bulk soils | 0.960 | 0.966 | 0.948 | 0.971 | 0.956 | 0.927 |

*All of the coefficients are significant at $P < 0.01$

enhanced by adding sulfate to the nutrient solution due to CdSO₄ complex (López-Chuken and Young 2010).

An opposite pattern was seen in +Cd treatments, where concentrations of Cd decreased in roots while Cd allocation in grains increased with increasing S application (Table 4, Figure 1). The results indicate that, when wheat was cultivated in highly contaminated paddy soil, S application tended to promote Cd transfer from root to aboveground tissues. This agrees with results reported by Zhao et al. (2003), who revealed that Cd concentrations of two spring wheat cultivars increased in shoots but tended to decrease in roots with the increase of K₂SO₄ fertiliser. There may be two reasons. Firstly, under high Cd stress, active oxygen species, e.g., superoxide anion (O₂⁻), hydroxyl radicals (·OH), and H₂O₂ are formed in wheat plant tissues (Ranieri et al. 2005, Khan et al. 2007). The Cd-induced active oxygen species can harm cell membranes and increase their permeability (Astolfi et al. 2005, Lin et al. 2007). Consequently, small-molecular-weight Cd-S compounds, such as Cd-cysteine and Cd-glutathione, can easily load into phloem tissues from xylem, and finally, deposit in grains. Cadmium-phytochelatins in the root cell vacuoles may release and decompose again due to the solubilisation of vacuole membranes. As a result, the limiting effects of S application present in -Cd treatments disappeared under high Cd stress. Secondly, the decrease of Cd allocation in roots with increasing S application (Figure 1) may be attributed to a decrease of root DWs by S application in +Cd treatments (Table 1).

Irrespective of S levels, Cd uptake by wheat was significantly ($P < 0.001$) higher in +Cd treatments than in -Cd treatments (Table 4). Besides, significant positive correlations ($P < 0.01$) existed between DTPA-Cd in rhizosphere/bulk soils and Cd concentrations in

plant tissues, as well as among tissues itself (Table 5). The results indicate that plant available Cd in paddy soils is the main source of Cd in wheat grains, which agree with results reported by Kikuchi et al. (2009). Our study clearly shows that S application could increase the accumulation of Cd in the grain of wheat, and hence, S fertilisation should be considered for winter wheat in Cd-polluted paddy soil. Some measures to reduce the bioavailability of Cd in the soil are recommended. For example, liming decreases the bioavailability of Cd in Cd-contaminated paddy soil (Zhu et al. 2016).

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REFERENCES

- Adhikari S., Ghosh S., Azahar I., Adhikari A., Shaw A.K., Konar S., Roy S., Hossain Z. (2018): Sulfate improves cadmium tolerance by limiting cadmium accumulation, modulation of sulfur metabolism and antioxidant defense system in maize. *Environmental and Experimental Botany*, 153: 143–162.
- Astolfi S., Zuchi S., Passera C. (2005): Effect of cadmium on H⁺ATPase activity of plasma membrane vesicles isolated from roots of different S-supplied maize (*Zea mays* L.) plants. *Plant Science*, 169: 361–368.
- Boussen S., Soubrand M., Bril H., Ouerfelli K., Abdeljaouad S. (2013): Transfer of lead, zinc, and cadmium from mine tailings to wheat (*Triticum aestivum*) in carbonated Mediterranean (Northern Tunisia) soils. *Geoderma*, 192: 227–236.

<https://doi.org/10.17221/558/2019-PSE>

- Cao Z.H., Meng C.F., Hu Z.Y. (2011): Sulfur in Agriculture and Environment of China. Beijing, Science Press, 27–31. (In Chinese)
- Chen H.Y., Teng Y.G., Lu S.J., Wang Y.Y., Wang J.S. (2015): Contamination features and health risk of soil heavy metals in China. *Science of The Total Environment*, S512–513: 143–153.
- Fan J.L., Hu Z.Y., Ziadi N., Xia X., Wu C.Y.H. (2010): Excessive sulfur supply reduces cadmium accumulation in brown rice (*Oryza sativa* L.). *Environmental Pollution*, 158: 409–415.
- Frolking S., Qiu J.J., Boles S., Xiao X.M., Liu J.Y., Zhuang Y.H., Li C.S., Qin X.G. (2002): Combining remote sensing and ground census data to develop new maps of the distribution of rice agriculture in China. *Global Biogeochemical Cycles*, 16: 1091.
- Gao Y., Ma M.Z., Yang T., Chen W.L., Yang T.T. (2018): Global atmospheric sulfur deposition and associated impact on nitrogen cycling in ecosystems. *Journal of Cleaner Production*, 195: 1–9.
- Khan M.I.R., Nazir F., Asgher M., Per T.S., Khan N.A. (2015): Selenium and sulfur influence ethylene formation and alleviate cadmium-induced oxidative stress by improving proline and glutathione production in wheat. *Journal of Plant Physiology*, 173: 9–18.
- Khan N.A., Singh S.G., Nazar R. (2007): Activities of antioxidative enzymes, sulphur assimilation, photosynthetic activity and growth of wheat (*Triticum aestivum*) cultivars differing in yield potential under cadmium stress. *Journal of Agronomy and Crop Science*, 193: 435–444.
- Kikuchi T., Okazaki M., Motobayashi T. (2009): Suppressive effect of magnesium oxide materials on cadmium accumulation in winter wheat grain cultivated in a cadmium-contaminated paddy field under annual rice-wheat rotational cultivation. *Journal of Hazardous Materials*, 168: 89–93.
- Lin R.Z., Wang X.R., Luo Y., Du W.C., Guo H.Y., Yin D.Q. (2007): Effects of soil cadmium on growth, oxidative stress and antioxidant system in wheat seedlings (*Triticum aestivum* L.). *Chemosphere*, 69: 89–98.
- López-Chuken U.J., Young S.D. (2010): Modelling sulphate-enhanced cadmium uptake by *Zea mays* from nutrient solution under conditions of constant free Cd²⁺ ion activity. *Journal of Environmental Sciences*, 22: 1080–1085.
- Lu R.K. (1999): *Methods of Agricultural Chemical Analysis in Soil*. Beijing, China Agricultural Science and Technology Publishing House, 198–200. (In Chinese)
- Matusiewicz H., Sturgeon R.E., Berman S.S. (1989): Trace element analysis of biological material following pressure digestion with nitric acid-hydrogen peroxide and microwave heating. *Journal of Analytical Atomic Spectrometry*, 4: 323–327.
- McLaughlin M.J., Lambrechts R.M., Smolders E., Smart M.K. (1998): Effects of sulfate on cadmium uptake by Swiss chard: II. Effects due to sulfate addition to soil. *Plant and Soil*, 202: 217–222.
- Mu T.T., Wu T.Z., Zhou T., Li Z., Ouyang Y.N., Jiang J.P., Zhu D., Hou J.Y., Wang Z.Y., Luo Y.M., Christie P., Wu L.H. (2019): Geographical variation in arsenic, cadmium, and lead of soils and rice in the major rice producing regions of China. *Science of The Total Environment*, 677: 373–381.
- Ranieri A., Castagna A., Scebba F., Careri M., Zagnoni I., Predieri G., Pagliari M., di Toppi L.S. (2005): Oxidative stress and phytochelatin characterisation in bread wheat exposed to cadmium excess. *Plant Physiology and Biochemistry*, 43: 45–54.
- Vityakon P., Meepech S., Cadisch G., Toomsan B. (2000): Soil organic matter and nitrogen transformation mediated by plant residues of different qualities in sandy acid upland and paddy soils. *NJAS – Wageningen Journal of Life Sciences*, 48: 75–90.
- Wieland G., Neumann R., Backhaus H. (2001): Variation of microbial communities in soil, rhizosphere, and rhizoplane in response to crop species, soil type, and crop development. *Applied and Environmental Microbiology*, 67: 5849–5854.
- Zhao Z.Q., Zhu Y.G., Li H.Y., Smith S.E., Smith F.A. (2003): Effects of forms and rates of potassium fertilizers on cadmium uptake by two cultivars of spring wheat (*Triticum aestivum*, L.). *Environment International*, 29: 973–978.
- Zhu H.H., Chen C., Xu C., Zhu Q.H., Huang D.Y. (2016): Effects of soil acidification and liming on the phytoavailability of cadmium in paddy soils of central subtropical China. *Environmental Pollution*, 219: 99–106.

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